

# Effects of irrigation with urban treated wastewater on the morphophysiology, accumulation of heavy metals and biochemical traits of *Casuarina glauca* Sieb

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**Abstract** - Waste water recycling is a strategy used to preserve fresh water resources, which are becoming increasingly scarce, particularly in Tunisia. Long-term reuse of treated waste water may have negative effects on plant growth and the environment, but short-term reuse may boost crop growth due to their high organic and mineral content. Irrigation of tree species by these unconventional waters during higher evapotranspiration periods for reforestation and biomass production could be a critical alternative for the recovery of these waters. The purpose of this research is to assess the effects of treated waste water (TWW) on the morpho-physiological behavior of *Casuarina glauca* over a period of 90 days. One-year-old seedlings were split into two groups, the first with drinking water (controls) and the second with treated wastewater (TWW). TWW had a positive effect on *C. glauca* plant growth, dry mass production, gas exchange, and the synthesis of total chlorophyll and primary and secondary metabolites. Furthermore, the findings demonstrated that this species has the ability to phytostabilize heavy metals, specifically Cu, Cd, and Pb.

Key words: Casuarina glauca, treated wastewater, gas exchanges, growth, heavy metals.

## 1. Introduction

Tunisia, like other Mediterranean North African countries, is experiencing a water scarcity, which poses a threat to long-term development (Slaimi et al. 2021). Thus, the annual renewable water resources of these countries in 2018 were 525m<sup>3</sup>/capita which just above thethresholdfor absolute water scarcity that is 500m<sup>3</sup>/capita (Frascari et al. 2018). Population growth, rising living standards, increased urbanization, desertification, overexploitation of groundwater, seawater intrusion into aquifers, deterioration of water quality and the intensification of the effects of climate change worsen this situation (Slaimi et al. 2021). Additionally, in the Mediterranean climatic zone such as North African country, rainfall was decreased as a result of climate change (Abouabdillah et al., 2010). Among the strategies that are included for the mobilization and development of water resources in Tunisia is the reuse of treatedwastewater in the agricultural sector where irrigation uses in Tunisia up to 80% of the water consumed (El Ayni et al. 2011). The total treated wastewatervolumewas estimated about 243 million m<sup>3</sup> witch present about 5% of mobilized water resources at the national level (Zouari et al. 2019). Their exploitation is still low because approximately 70% of their volume is removed in the sea and valleys, and the rest is used for various sectors, including agriculture, irrigation of green areas, golf courses, and groundwater recharge (Zouari et al. 2020). The use of TWW would be a strategy for crop production due to its high content of nutrients essential for plant growth. However, high mineral salts, heavy metals and organic compounds in this kind of water can deteriorate soils, crops and human health (Shehzadi et al.2014). As a result, its use for irrigation must be done with caution to mitigate its negative effects (Slaimi et al. 2021). Irrigation with TWW for the production of forage or food crops in lands with limited freshwater supplies is a well-known solution (Cornel and Weber 2004; Raes et al. 2009). In addition its use as irrigation supply for afforestation of marginal lands could support the rehabilitation of such degraded area which represents around 75 % of the Tunisian's land area (Béjaoui et al. 2008). Failures of tree or shrub plantations are mostly related to the low survival rate of young plants due to the lack of water after transplantation (Béjaoui et al. 2008), irrigation with treated wastewater could enhance the successof these plantations by providing water supply for seedlings, particularly during the early phase of growth. In addition, afforestation of degraded soils could contribute to the conservation of



water and soil, the fight against global warming through carbon sequestration and mitigation of greenhouse gas emissions. Also, It could improve water quality by filtration and absorption of harmful chemicals, including metals, organic compounds such as fuels by transforming them to less harmful substances stored in the roots, stems or leaves (Kuzovkina and Quigley, 2005; Tan and Yeo, 2009). One of the most solutions for the restoration of degraded soils is the use of woody shrubs or trees, highly resilient to harsh environmental conditions (Diem and Dommergues, 1990). The use of trees for phytoremediation and forest creation becomes increasingly essentialaslow-cost ecological strategies to rehabilitate and improve ecosystem benefits (e.g., timber production, firewood, and fodder) (Ghazouani et al. 2020)and to mitigate climate change effects over the next decades (El Moussaoui et al., 2019). Among the woody species witch able to establish the degraded soil figure the *Casuarina* species. They have a fast growth and cangrow up to 3m per year to reach a final height of 30 m (Ghazouani et al. 2020). They also present an economic and ecological impact through the use of their wood for the construction of the production of plywood (Zhong et al. 2010). These Australian origin species not widespread in the Mediterranean basin are known by their high tolerance of various environmental constraints (Ghazouani et al. 2020). Several studies have shown its potential for adaptation to various conditions of environmental stress such as flooding, drought, salinity, alkalinity, acidity, and mineral or organic pollution (Béjaoui et al. 2008; Bargali, 2011; Ghazouani et al. 2020). Historically, these woody plants were planted on poor and contaminated soils in order to increase their fertility (Duhoux and Franche, 2003). Casuarina trees are frequently used for windbreaks, erosion control, dunes stabilization in particular in lands where the water resources are limited (Albouchi et al. 1989). Casuarinaceae family form symbiotic N-fixing associations with actinomycetes (Frankia) as well as ectomycorrhizal and endomycorrhizal fungi. (Diem 1996). This symbiotic enhances the resistance and root plasticity system of *Casuarina* species in harsh conditions or degraded soils (Duhoux and Franche, 2003). Casuarina glauca was the candidate species for this study due to its wide using in agroforestry, for woody

*Casuarina glauca* was the candidate species for this study due to its wide using in agroforestry, for woody production programs and soil rehabilitation (Zhong et al. 2013, Slaimi et al. 2021). In order to ensure the tolerance of *C. glauca* used for afforestation, it would be necessary to evaluate their behavior after their irrigation with treated wastewater. The aim of this study is (i) to evaluate the treated wastewater effects of on *C. glauca* morpho-physiological behavior after 90 days (ii) to determine to accumulate and translocate heavy metals.

#### 2. Materials and methods

#### 2.1. Experimental setup and culture Conditions

The study was carried out at the nursery of the National Institute of Research in Rural Engineering, Waters and Forests of Tunisia (INRGREF) in a semi-arid bioclimate (36 °50 N, 10 ° 14 E and 3 m above sea level). The annual average of rainfall was around 475 mm and the monthly average of temperatures varied between 7.2 °C and 34.8°C. The experiment was conducted under semi-controlled conditions using one year old C. glauca plants. Homogeneous seedlings with the same height growth have been put in plastic pots of 50 cm in diameter and 60 cm in depth, containing 10kg of substrate consisting of a mixture of sand and loam (30% - 70%). They were divided into two batches, the first one was irrigated with tap water (control C) and the second with urban secondary treated waste water (TWW) collected from the outlet of wastewater treatment plant Kalâat El Andalous of Tunisia (N: 36° 53' 51".55 E: 10° 12' 01".71). This is an ecological wastewater treatment plant that provides secondary effluents treated with lagoons. The main physicochemical and microbiological characteristics of TWW are shown in table 1. The characteristics of the tap water used for the control are (61 mg/l of Ca<sup>2+,</sup> 23 mg/l of Mg<sup>2+,</sup> 15.1 mg/l K+, 0.569 mg/l of PO43, 314.11 mg/l of Cl and 0.066 mg/l Fe3+). The plants in both treatments were irrigated every 2 days according to the water holding capacity of the soil to prevent water leakage from pots throughout the experiment, which lasted 90 days. At the end of this period (90 days), several morphological, physiological, biochemical parameters and concentration of heavy metals in plant tissues have been measured. The adopted experimental design type was a random full. Each plant that is represented individually by a pot has been considered as a repetition in a completely randomized design which contained 20 repetitions per treatment.



**Table1**: Microbiological and physicochemical properties of treated wastewater (TWW) used according to the standards Tunisian for the discharge of effluents into hydraulic natural flows (NT 106.02).

Parameters	Units	TWW	NT 106.02	
Total coliforms	UFC per 100ml	4. $3 \times 10^3$	-	
Fecalcoliforms	UFC per 100 ml	3. $5 \times 10^3$	2000	
Fecal streptococcus	UFC per 100ml	2. $5 \times 10^5$	1000	
pH		$7.94 \pm 0.01$	6.50-8.50	
Conductivity	mS/cm	5.32±0.02	7	
Salinity	g/l	$3.8 \pm 0.2$	-	
Temperature	°C	$20 \pm 0.6$	25	
BOD <sub>5</sub>	mg O <sub>2</sub> /l	$37 \pm 0.03$	30	
COD	mg/l	385 ±0.11	90	
Suspendedmatter	mg/l	35±0.04	30	
Nitrogen	mg/l	$22.58 \pm 0.17$	30	
Phosphorus	mg/l	$2.38 \pm 0.19$	0.05	
Ortho phosphate(PO <sub>4</sub> <sup>3-</sup> )	mg/l	$18.396 \pm 0.53$	0.05	
Magnesium	mg/l	139.9±0.005	200	
Sodium	mg/l	$1192 \pm 0.01$	300	
Calcium	mg/l	$338 \pm 0.02$	500	
Sulfate	mg/l	$750 \pm 0.01$	100	
Potassium	mg/l	166. 8 ±0.04	50	
Chloride	mg/l	$1532 \pm 0.03$	600	
Cadmium	mg/l	0. 2145 ±0.01	0.005	
Manganese	mg/l	$0.16597 \pm 0.09$	0.5	
Copper	mg/l	$0.1181 \pm 0.007$	0.5	
Lead	mg/l	$0.3114 \pm 0.12$	0.1	
Nickel	mg/l	$0.531 \pm 0.005$	0.2	
Zinc	mg/l	$0.9043 \pm 0.02$	5	

## 2.2. Growth and biomass of plants

The growth of the *C. glauca* plants irrigated with TWW and fresh water hasbeen assessed by the measurement of the height of the main stem and weighed shoots and roots biomass. In order to evaluate the effect of TWW treatmenton the growth of seedlings, the relative growth rates (RGR) of plantheight was calculated using 10 plants per treatment according to the equation described by Salim and Pitman (1983): RGR= (Ln W2-Ln W1)/ (t<sub>2</sub>-t<sub>1</sub>)

Where W2= Height of plant (cm) at time  $t_2$ 

 $W1 = Height of plant (cm) at time t_1$ 

t<sub>1</sub> and t<sub>2</sub>(separatedby days depending on doubling time)

At the end of the experiment, the dry mass produced in the aerial and theroot part was measured on 10 plants per treatment. The roots were extracted after sievingof the substrate using a mesh of 2 mm. The biomass of plants was determined by weighing after dryingat 70°C for 48 hours.

#### 2.3. Measurements of gas exchange, chlorophyll content and relative water content

Net photosynthesis (A), stomatal conductance(gs), transpiration (E)and intercellular CO2 concentration(Ci)were determined between 09:15 and 11:15 am usingan infrared analyzer the LCPRO + (LCpro + model, ADC Bio Scientific LTD., Hoddesdon, England) on five seedling selected randomly by treatment, under the following conditions: an active photosynthetic flux of photons (PPFD) of 980  $\pm$  64µmol.mol<sup>-1</sup>m<sup>-2</sup>s<sup>-1</sup>, a CO<sub>2</sub> concentration of350 µmol.mol<sup>-1</sup> and a foliar temperature of 30  $\pm$  2 ° C.The intrinsic water use efficiency (iWUE)has been calculated as the ratio of net photosynthesis to the stomatal conductance(Farquhar et al. 1982).

For determining chlorophyll concentration concentrations, 100 mg of fresh needle was crushed in 10 ml of 80 percent acetone. The samples were incubated at 4°C overnight before being centrifuged at 10,000g for 10 minutes. The extract was collected, and the supernatant's absorbance was measured at 645 and 663 nm with a spectrophotometer (Shimadzu UV-1800 PC model Kyoto, Japan). The total chlorophyll content was calculated using the equation of Arnon (1949) and expressed as mg. g-1 dry mass.



The water status of the plants for each treatment was estimated through the measurement of the relative water contents of treated and control leaves. This parameter was measured on ten needlesused by the same plants used for the gas exchange and chlorophyll measurements. The relative water contents (RWC) was calculated according to the formula:

 $RWC = (P_f - P_s)/(P_{turg} - P_s) \times 100$ 

# With:

 $\mathbf{P}_{f}$ : the fresh mass determined immediately after excision of shoots.

 $P_{turg}$ : the saturation mass determined after immersion of the sectioned petiole in distilled water and its passage to the dark at 25°C for 24 h.

 $P_s$ : the dry mass determined after drying at 70°C for 48 hours.

# 2.4. Determination of soluble sugars and proline

## 2.4.1. Soluble sugars extraction

The extraction of soluble sugars was performed according to the technique described by Albouchi (1997). These measures have been carried out on five samples of mature needles which were randomly selected for each treatment. For each repetition, 100 mg of leaf dry matter finely crushed were mixed with 10 ml of 80% ethanol. Themixture was heated in a water bath at 70  $^{\circ}$  C for 30 minutes and then centrifuged at 10 000 g for 15 minutes. The soluble sugars in the supernatant were assayed by 0.2% anthrone in the presence of 80% ethanol. The optical density was measured at 640 nm in a spectrophotometer (Lambda 40, Perkin Elmer, USA). The concentrations of soluble sugars determined using glucose standard and expressed as mg g<sup>-1</sup>dry matter.

## 2.4.2. Proline extraction

Proline was extracted according to the method described by Monneveux and Nemmar (1986), using Lproline as a standard. One hundred mg of needles dry powder were placed in a test tube, to which 2 ml of methanol (40%) have been added. The mixture was heated in a water bath at 85 ° C for 60 minutes. After cooling, 1 ml of the extract was tacked and mixedwith 1 ml of acetic acid, 25 mg of ninhydrinand 1 ml of a mixture containing (120 ml of distilled water, 30 ml of acetic acid and 80 ml of orthophosphoric acid). The tubes were thenincubated in a water bath at 100 C for 30 min. After cooling, 5 ml of toluene were added and taken for centrifugation. The supernatant containing the proline was recovered and dried by addition of Na<sub>2</sub>SO<sub>4</sub>. The optical density was measured at 528 nm in a spectrophotometer (Lambda 40, Perkin Elmer, USA). The concentrations of proline are expressed as  $\mu$ mol g<sup>-1</sup> of dry matter.

#### 2.5. Determination of phenolic compounds

The extraction of phenolic compounds was performed according to the method of Mau et al. (2001), by mixing 2.5 g of needles or roots dried matter with 25 ml of distilled water. The mixture was agitated for 30 min and kept at rest for 24 hours at 4°C in darkness. Finally, the mixture is filtered on paper Wattman. The obtained extracts were stored at 4°C for later use in the dosage of total polyphenols and flavonoids.

# 2.5.1. Determination of total polyphenols

A volume of 125  $\mu$ l of the plant extract diluted 10 times was mixed with 500 $\mu$ l of distilled water and 125 $\mu$ l of Folin-Ciocalteureagent. After vigorous agitation of the mixture followed by a rest 3min, a volume of 1250  $\mu$ l of CO<sub>3</sub>(Na)<sub>2</sub> (7 %) was added. Finally the mixture was adjusted by distilled water to 3 ml. After resting for 90 min in the dark, reading the absorbance was performed using a spectrophotometer (Lambda 40, Perkin Elmer, USA) at 760 nm. The standard range was prepared with gallic acid at concentrations ranging from 50 to 500 mg.L<sup>-1</sup>. The levels of polyphenols wereexpressed as mg of gallic acid equivalent per gram of dry matter (mg EAG.g<sup>-1</sup> MS). (Bettaib Rebey et al, 2016).

# 2.5.2. Determination of total flavonoids

A volume of 0.25 ml of extract was diluted for 5 times with 0.075 ml of  $NaNO_2$  (5%). The mixture was standing for 6 min in temperature room before adding 0.15 ml of aluminum chloride (AlC<sub>13</sub>, 10%), freshly



prepared. A second incubation for 5 min at room temperature was performed and also followed by the addition of 0.5 ml of NaOH (1M). The mixture is subsequently adjusted with distilled water to a final volume of 2.5 ml. The absorbance reading was made at 510 nm. The standard range was prepared with catechin at increasing concentrations ranging from 50 to 500 mg.L<sup>-1</sup>. The concentrations of flavonoids were expressed as mg equivalent of catechin per gram of dry matter (mg EC.g<sup>-1</sup> MS).

## 2.6. Heavy metals analyses

The assay was performed on shoots and roots samples of five plants randomly selected for each treatment previously rinsed with distilled water and then dried at 80 °C for 72 h. The dried samples were ground and digested in the presence of three acids solution (HNO<sub>3</sub>:HClO<sub>4</sub>: H<sub>2</sub>SO<sub>4</sub>; 10: 4: 1) on a hot plate at 300 °C for 2 h. The extracts, were also filtered on filter paper (Whatman No. 1) without ash. The heavy metals concentrations in filtrate were determined by using atomic absorption spectrophotometry (Model COULD 9100, Philips, Cambridge,UK). The concentrations of heavy metals were expressed as mg.kg<sup>-1</sup> · dry mass (DM).

The metals transfer capacity of *C. glauca* from roots to shoots was evaluated at the end of the experiment with the calculation of translocation factor (TF) (Mackay and Fraser 2000) according to the following formula:

*TF* = *Metal concentration in shoot/Metal concentration in root* 

## 2.7. Statistical analysis

The data were subjected to statistical analysis using the software SPSS, 16.0 (SPSS Institute Inc., 252 Chicago, Illinois, USA). The comparison of the averages for the different measured variables was carried out using the test Student at the threshold of 5%.

#### 3. Results

#### 3.1. Growth, biomass production and gas exchange of seedlings.

After 30 days, the RGR of plants height (RGRh) was similar in control and treated plants which areabout 0.006 cm.d<sup>-1</sup> (Figure 1). The RGRh of treated plants increased by 20% and 24% respectively after 60 and 90 days of irrigation with TWW compared to control. Then, after 90 days of irrigation, shoots and roots dry mass of treated plants increased by 26% and 24%, respectively, compared to control plants. Also, in treated plants, net assimilation (A), stomatal conductance (gs), transpiration(E) and CO2 intercellular (Ci) increased significantly (p < 0.05) by 57%, 24%, 68% and 22%, respectively compared to control plants (Figure 2). Furthermore, net assimilation (A) oftreated plantswas 2.4 times greater than their stomatal conductance (gs). Similarly, treated plants have higher intrinsic water use efficiency (WUEi) than control plants (Figure 2).



**Figure 1.** Relative growth rate of plant height and biomass production of shoots and roots dry mass of *Casuarina glauca* irrigated with tap water (control) and secondary treated wastewater (TWW) after 90days. For each parameter, the bars represent the mean values  $\pm$  SE (n = 5) associated with different letters (a,b) indicate significant differences among the two treatments. The confidence intervals were calculated at5 % threshold.





**Figure 2.** Net photosynthesis rate (A), stomatal conductance (gs), transpiration (E), intracellular CO2 (Ci), Intrinsic Water Use efficiency (WUEi), total chlorophyll and relative water of *Casuarina glauca* irrigated during 90 days with tap water (control) and secondary treated wastewater (TWW). For each parameter, the bars represent the mean values  $\pm$  SE (n = 5)associated with different letters (a,b) indicate significant differences among the two treatments. The confidence intervals were calculated at 5 % threshold.

# 3.2. Water content and biochemical Changes in *C. glauca*

The relative water content (RWC) of treated plants has not significantly changed after 90 days compared to controls indicating adequate tissue hydration (Figure 2). Total chlorophyll content was also significantly



enhanced (p< 0.05) after irrigation with TWW during 90 days, it was higher 40% in treated than in control plants of *C. glauca* (Figure 3). The phenolic compounds of plants irrigated with TWW were significantly greater (p< 0.05) than in *C. glauca* control plants (Figure 3). In addition, total polyphenols and flavonoids in shoots of treated plants were 2.3-fold and 1.5-fold higher than in control plants. However, the soluble sugars in plants irrigated with TWW were significantly higher than in control plants after 90 days. In contrast, there was no significant difference of proline concentration between treated and control plants *C. glauca*.



**Figure 3.** Proline content, soluble sugar, polyphenol and flavonoids content of *Casuarina glauca* irrigated during 90 days with tap water (control) and secondary treated wastewater (TWW). For each parameter, the bars represent the mean values  $\pm$  SE (n = 5) associated with different letters (a,b,c,d) indicate significant differences among the two treatments. The confidence intervals were calculated at 5 % threshold.

#### **3.3.** Effect of treated wastewater on accumulation and translocation of heavy metals

The results showed higher heavy metal accumulation in roots than in shoots of *C. glauca* irrigated with, such as Cu, Zn, Cd, Ni, and Pb, except for Mn that is more accumulated in shoots (Figure 4). Their concentrations in roots were (3,2; 1,8; 4; 3.5 and 3.5-fold) higher than in shoots, while the Mn concentration in shoots was 3 times higher than in roots. The irrigation with TWWduring90 days has significantly (P <0.05) increased the concentration of these metals in plants of *C. glauca*. Cd is accumulated in higher proportion in the two parts of treated plants. It was 4 times higher in shoots and 8.8 times higher in roots compared to control plants. The translocation factor (TF) of heavy metals in *C. glauca* is shown in table 2. The highest TF (> 1) was recorded for Mn, related to the preferential translocation of this element to the leaves. The lowest TF (0,12) has also been noted for the Cd. Furthermore, the TF of Zn, Cu, Ni, and Pb ranged from 0.56 to 0.28.

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**Figure 4.** Concentrations of zinc (Zn), manganese (Mn), nickel (Ni), cupper (Cu), cadmium (Cd), lead (Pb), in shoots and roots of *Casuarina glauca* plants irrigated during 90 days with tap water (control) and secondary treated wastewater (TWW). For each parameter, the bars represent the mean values  $\pm$  SE (n = 5) associated with different letters (a,b,c) indicate significant differences among the two treatments. The confidence intervals were calculated at 5 % threshold.

**Table 2:** Translocation factor (TF) of heavy metals in *Casuarina glauca* irrigated with secondary urban treated wastewater after 90 days.

Métaux lourds	Zn	Mn	Cu	Ni	Pb	Cd
TF	$0.56{\pm}0.02~b$	3.19± 0.01 a	$0.35 \pm 0.03 \text{ c}$	$0.32 \pm 0.03 \text{ c}$	$0.28 \pm 0.02 \text{ d}$	0.12± 0.01 e

The averages followed by the same letter are not significantly different according to the Newman and Keuls test (p < 0.05).

#### 4. Discussion

Irrigation with urban secondary TWW for 90 days significantly increased the growth of C. glauca compared to control plants which were irrigated with fresh water. The TWW increased the height of the RGR of treated plants (fig.1). The same effects of irrigation with TWW have been demonstrated on the biomass production



of treated plants (fig.1). Thus, the TWW stimulated biomass production in the aerial parts as well as roots (fig.1). This growth was related to the nutrient content in TWW, such as mineral nitrogen, phosphorus, potassium, and micronutrients (Rais et al. 2009). These elements, which are essential for growth, stimulated the biomass productivity of C. glauca (Karnosky, 2003). Taking into account the fertilizer supply of phosphorus, nitrogen, potassium, and macronutrients, some authors believe that TWW could play the same role as fertilizer, depending on the concentration of these elements in water, the type of crop, and the level of soil fertility (Singh and Agrawal, 2010; Justin et al. 2010). The results also indicate that irrigation with TWW for 90 days stimulated the production of biomass without any negative effects of pollutants. These findings are consistent with Kumar and Reddy's (2010) findings of positive growth in Casuarina equisetifolia trees irrigated with TWW for 13 months. In contrast, Zarati et al. (2015) reported that the irrigation with TWW for 60 days reduced the growth and thephotosynthesis rate of *P. nigra*. The positive responses of physiological characteristics of plants irrigated by treated wastewater with high levels of sodium and chloride, which are respectively 4 and 2.6 times higher than the norm (Table 1), suggest that C. glauca seedlings tolerate the salinity of waterwhich is also confirmed byScotti-Campos et al, (2016). However, no significant differences (p > 0.05) of relative water contents (RWC) were observed between treated and control plants, reflecting an osmotic adjustments to prevent oxidative damage (Hachani et al. 2021). Total chlorophyll was higher in plants irrigated withTWW. The stimulation of chlorophyll production could berelated to the preferential compartmentalization of Mn in leaf cell organelles (Emamverdian et al. 2015). The Mn has several metabolic roles in different cellular compartments that induce plant growth (Emamverdian et al. 2015). It is an essential cofactor of several biochemical reactions in plants. Moreover, the most metabolically dependent plant on Mn is the water-splitting reaction in photosystem II (Alejandro et al. 2020). In contrast to the present study, several studies reported that heavy metal accumulation decreased chlorophyll production (Padmaja et al. 1990; Mysliwa-Kurdzieland Strzalka, 2002). Moreover, the higher concentrations of heavy metals such as Cd and Pb in the needles of treated plants did not affect the chlorophyll synthesis, which indicates the metal stress tolerance of C. glauca during the experimental time. The higher chlorophyll concentration of treated plants increased the photosynthetic rate and, indirectly, the growth and biomass production. Similarly, other studies reported that irrigation with TWW increased the photosynthetic rate of plants such as *Populus nigra*, *Salix nigra*, and *Populus canadensis* (Borghi et al. 2007; Singh and Agrawal, 2010; Zarati et al., 2015). In addition, the higher availability of nitrogen (N) in TWW has stimulated the photosynthetic performance of plants. This stimulation is linked to the involvement of N in the production of essential molecules for the photosynthetic system such as chlorophyll and chloroplast composition (Deroche, 1983; Bassi et al. 2018). Yet, another study showed a positive correlation between nitrogen supply and the increase in photosynthetic rate in sugarcane (Bassi et al. 2018). The higher stomatal conductance (gs) of plants irrigated with TWW as compared to ground water irrigated ones (Fig. 3) could be due to potassium availability in TWW (Table 1). Indeed, potassium is one of the main osmotically active solutes involved in the stomatal opening which leads to the increase in  $CO_2$  uptake and photosynthesis (Dietrich et al. 2001). In addition, a slight increase of WUEi was noted in plants irrigated with TWW that indicates plants tolerance to wastewater pollutants. The higher accumulation of solutes such as anions, nutrient compounds in treated plants than in control plants can prevent plants against a greater water relations reduction by regulating cellular osmotic potential and stomatal closure (Gómez-Bellotet al. 2020). In fact, water is the exchange unit that is required for the acquisition of  $CO_2$  for plants. Thus, the WUEi of treated plants reflects a lower water amount spent per  $CO_2$  molecule assimilated (Farooq et al., 2009). In the current experiment, primary metabolites such as proline and soluble sugar content increased in plants after irrigation with TWW. The high intracellular levels of accumulation of these osmoprotective compounds regulates the osmotic balance of cells and preserves the subcellular structures for better water absorption under stress (Zhang et al. 2010; Sharma et al. 2012). However, the accumulation of proline is a possible mechanism to detoxify plants through ROS scavenging against oxidative damage to mitigate the metallic stress effects (Hassan et al. 2018). Irrigation with TWW has increased specific phenolic compounds in plants, such as total polyphenols and flavonoids. These secondary metabolites have several physiological functions, such as antioxidant properties such as the protection of cells from potential damage related to heavy metal accumulation (Agati et al. 2012; Zhang et al. 2015). Higher accumulation of heavy metals (Zn,



Cu, Ci, Pb and Cd) in roots of C. glauca plants than in shoots has been observed. The accumulation of metals differs between organs and tissues of the same plant (Rotkittikhun et al. 2006). The heavy metals amount was significantly enhanced in root for Cu, Ni, Pb, and in particular for Cd which increased 8.8-fold compared to in control plants. These amounts indicate that C. glauca is non-hyper accumulative species because all heavy metals concentrations were less than 1000 mg.kg<sup>-1</sup>(Baker and Brooks, 1989). Hyper accumulative plant species have a strong capacity to extract metals by transferring them to aerial part, but the tolerant plants tend to restrict this transfer to soil-root or root-shoot system (Yoon et al. 2006). The ability of plants to transfer the metals from the root to the aerial part is determined by the translocation factor (TF). Thus TF < 1indicates a low heavy metals transfer from roots to the aerial part suggesting that these plants accumulate the metals in their roots or rhizomes more than in their aerial part (Usman et al, 2019). This case applies to C. galuca plants which have shown an FT < 1 for the elements (Zn, Cu, Ni, Pb, and Cd). In addition, the very low translocation of the Pb (TF = 0,28) and, in particular, Cd (TF = 0,12) indicates that the C. glauca plants would have limited the transfer of these elements from the roots to the leaves in order to decrease their toxicity. These two elements are not essential for the development of the plant, their accumulation could be toxic to photosynthetic activity, the synthesis of chlorophyll and the activities of antioxidant (Slaimi et al. 2021; Touati et al. 2019). The sequestration of these metals in the roots, which limited their toxic effects on leaves, would be an adaptive mechanism of C. glauca to mitigate the metal stress. This mechanism could also be a metal-exclusive character in C. glauca (Mganga et al. 2011; Ghazouani et al. 2020). In addition, our study showed the phytostabilisation capacity of C. glauca for these metals in order to reduce their mobility and their diffusion in the soil (Susarla et al. 2002). It reduces the mobility of metals and their leaching into groundwater which limits their bioavailability and their passage into food chains (Yoon et al, 2006). The stabilization and the extraction of heavy metals by plants are the result of the ability of root exudates to change physicochemical characteristic of the rhizosphere. Moreover, these exudates are able to modify the soil pH which increased the chelating, the solubilization and the uptake of heavy metals and micronutrients (Shen et al. 2017). Moreover, the strong C. glauca tolerance of Cd high level in roots may be due to the richness of root exudates with secondary metabolites with low molecular weight. These metabolites are capable to create a Cd-phytochelatin complex at the soil-root interface which allowing the plants tolerate the Cd toxicity (Bali et al. 2020). This mechanism could provide efficient conditions for natural removal or stabilization of pollutants from soil (Yoon et al. 2006).

#### 5. Conclusion

The current study investigated the fertilizing effect of irrigation with urban treated wastewater on *C. glauca* seedlings over a 90-day period. This was demonstrated by a faster increase in height growth as well as an increase in biomass production. C. glauca was recommended for the rehabilitation of marginal and degraded lands due to its rapid growth, higher biomass production, adaptation to extreme conditions, and phytostabilization potential of heavy metals. The use of TWW as a source of water supply to ensure the success of the early phase of plantation could be an advantage of this process. However, on-site planting testing is required to evaluate this species in natural conditions and over a longer period of time. Previous studies have shown the effectiveness of the symbiosis *C. glauca*—Frankia in the bio-remediation of soils polluted with industrial effluents. The interaction between *C. glauca* and mycorrhizae, which could affect the uptake and translocation of metals, deserves further investigation. The use of TWW as a source of water supply to ensure the satural conditions and over a longer period of time planting testing is required. Previous function between *C. glauca*—Frankia in the bio-remediation of soils polluted with industrial effluents. The interaction between a dvantage of this process. However, in order to evaluate this species in natural conditions and over a longer period of time, on-site planting testing is required. Previous research has demonstrated the efficacy of the *C. glauca*—Frankia symbiosis in the bio-remediation of soils polluted with industrial effluents. The interaction between *C. glauca*—Frankia symbiosis in the bio-remediation of soils polluted with industrial effluents. The interaction between *C. glauca*—Frankia symbiosis in the bio-remediation of soils polluted with industrial effluents. The interaction between *C. glauca*—Frankia symbiosis in the bio-remediation of soils polluted with industrial effluents. The interaction between *C. glauca* and mycorrhizae, which

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